Extinction-colonization dynamics structure genetic variation of spotted sunfish (*Lepomis punctatus*) in the Florida Everglades

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Abstract

The population genetics of aquatic animals in the Florida Everglades may be strongly influenced by extinction and colonization dynamics. We combined analyses of allozyme and microsatellite loci to test the hypothesis that two levels of population structure are present for spotted sunfish (Pisces: Centrarchidae: Lepomis punctatus) inhabiting the Everglades. We hypothesized that annual cycles of marsh dry-down increase local-scale genetic variation through a process of local extinction and colonization; we hypothesized that barriers to gene flow by levee/canal systems create a second, regional level of genetic variation. In 1996 and 1997, we sampled spotted sunfish from 11 Everglades sites that were distributed in three regions separated by levees. We documented patterns of genetic variation at 7 polymorphic allozyme loci and 5 polymorphic microsatellite loci. Most genetic variation was present among local populations, according to both types of genetic markers. Furthermore, samples from marsh sites were heterogeneous, while those from canals were not. These data supported our hypothesis that dry-down events and local population dynamics in the marsh have a significant effect on population genetic structure of spotted sunfish. We found no support for our hypothesis that water-management structures superimpose a second level of genetic structure on this species, possibly because canals obscure historical structure by facilitating gene flow or because the complete canal system has been in place for fewer than 20 generations of this species. Our data suggests a continent-island (canalmarsh) structure of populations with high gene flow among regions and recurrent mixing in marshes from canal and creek habitats.

Keywords: allozymes, extinction, Florida Everglades, *Lepomis punctatus*, microsatellites, migration, population structure

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Introduction

The genetic structure of a metapopulation may be shaped by the geographical scale and pattern of dispersal leading to colonization following local extinction events (Wade & McCauley 1988; McCauley *et al.* 1995). For example, colonists from a small number of source sites elevate the level of genetic variation among populations within a metapopulation (propagule-pool colonization). In contrast, when colonists from many sources re-populate a site, local extinction leaves little impact on metapopulation genetic

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structure (migrant-pool colonization). *F*-statistics (F_{ST}) can be used to distinguish stable entities of a population, identify newly colonized sites and indicate the dominant mode of colonization (Slatkin 1977; Wade & McCauley 1988; Whitlock & McCauley 1990). The genetic structure of a metapopulation and the dominant mode of colonization for a species may depend on the spatial scale and physical structure of the environment being considered.

The effects of habitat fragmentation on population genetic structure can vary widely and may even be concealed or diminished by population-level genealogical relationships that persist when gene flow no longer occurs between two populations. Pannell & Charlesworth (1999) noted that $F_{\rm ST}$ is stable to fluctuations in the number of

occupied sites, such as in a population that is expanding and contracting over time. They cautioned that F_{ST} is relatively insensitive to the mode of colonization if the rate of extinction is less than or equal to the rate of gene flow between populations. Further, F_{ST} may be a poor estimate of population divergence because it is sensitive to the total diversity (the denominator) unrelated to the difference between populations. Thus *F*-statistics may underestimate genetic divergence between populations, especially when allelic diversity is high (Nagylaki 1998; Pannell & Charlesworth 1999).

The Everglades can be characterized as a south flowing river with a braided structure of relatively deep sloughs separated by densely vegetated shallow ridges, and bound by relatively short-hydroperiod wetlands. The slough habitats are routinely fragmented by seasonal dry-down events that temporarily reduce and/or eliminate connections between them (separated by hundreds of metres to kilometres) and concentrate aquatic organisms into deepwater refuges (Loftus & Kushlan 1987; Trexler et al. 2001). Starting before 1900 and culminating in the late 1960s (reviewed in Blake 1980; Light & Dineen 1994), canals that may enhance fish dispersal and levees that may limit or preclude migration by fish have subdivided the ecosystem into regional management units covering hundreds of square kilometres. These two forms of habitat fragmentation, seasonal dry-down events and structural fragmentation, differ in their implications for population genetic structure. Seasonal dry-downs may cause the local extinction of populations, and colonists may re-populate local habitat from a variety of sources. Additionally, dry-downs may encourage mixing by relatively long-range movement of fishes into refuge habitats. Movement of individuals into deep-water refuges from a diversity of marsh sites during a dry-down event may result in population mixing revealed genetically by deviation from Hardy-Weinberg equilibria and deficiency of heterozygotes (Wahlund effect: Hartl & Clark 1997). Structural fragmentation by canals may create corridors that facilitate directional movement, while levee barriers may reduce the dimensionality of movement through the habitat. Thus, structural fragmentation may create either corridors that homogenize populations, or barriers to gene flow facilitating genetic differentiation of subpopulations on either side.

We have used *F*-statistics to identify the metapopulation genetic structure and distinguish between two modes of colonization (propagule and migrant pool) for a common centrarchid, the spotted sunfish (*Lepomis punctatus*), found throughout the Florida Everglades. We combined analyses of allozyme and microsatellite loci to test the hypothesis that two levels of population structure may be present for spotted sunfish in the Everglades: one of local variation resulting from extinction and colonization; and a second of regional variation resulting from barriers to gene flow by levees and canals superimposed over the historical pattern of a large unimpeded river. If extinction and colonization are important, relatively greater F-statistics will be observed between sites that are subject to extinction-colonization dynamics compared to sites that are stable. Further, if colonization occurs from many different sources, including some distant from the receiving population (migrant-pool colonization), then local extinctions will have little effect on the population genetic structure within water management units. In the latter scenario, relatively large F-statistics among regions remain plausible. In this case an island model, perhaps at the scale of management units, would be appropriate to describe genetic diversity and the greatest Fstatistic would be expected at the regional scale. The greatest F-statistic would be recorded at the local scale within water management units if colonization were from a small number of local sites (propagule-pool colonization) (Slatkin 1977; Wade & McCauley 1988).

Materials and methods

Sample collection

In the Everglades ecosystem, spotted sunfish are relatively large (85 mm adult mean standard length [SL = length from the tip of the snout to the base of the caudal peduncle]), long-lived fish (sexual maturity at approximately 1 year); most Everglades fish are cyprinodontiformes with approximately annual life cycles (Loftus & Kushlan 1987; Trexler et al. 2001). We collected 278 spotted sunfish between 1996 and 1997 from eight canal and three marsh-pond sites distributed throughout Shark River Slough in Everglades National Park (ENP) and Water Conservation Areas 3A (WCA-3A) and 3B (WCA-3B) (Fig. 1; Table 1). Canal and levee constructions separate these three regions of the Everglades, with WCA-3B being completely isolated hydrologically from the rest of the ecosystem by a system of levees completed in 1963. Gates permit a regulated flow of water southward from WCA-3A to ENP. The sample size from each site ranged from 12 to 52, and the fish were obtained by angling and electrofishing. Upon collection, animals were transferred to the laboratory, standard length was measured, and they were stored at -80 °C prior to genetic analyses.

Sampling was conducted in the dry season of 1996 to obtain specimens from areas where they were concentrated because of declining water levels. However, an insufficient number of specimens was collected, so additional samples were obtained in the dry season of 1997 and pooled into a single sample with the fish from the previous year. We considered that pooling between the 2 years was appropriate because spotted sunfish live several years, permitting the adult fish from both years to be treated as members of the same cohort. Also, there was no evidence

Site	Habitat	Region	Latitude	Longitude	Ν	SL
L1	marsh	WCA-3A	25.51.67	80.43.49	23	84.5 (5.85)
L3	canal	WCA-3A	25.59.45	80.50.35	25	92.7 (2.51)
L11	canal	WCA-3A	25.45.70	80.49.40	25	100.0 (4.03)
L21	canal	WCA-3A	26.06.54	80.36.26	14	93.2 (2.98)
L23	canal	WCA-3A	26.03.80	80.26.30	25	65.3 (1.74)
L25	canal	WCA-3A	26.08.82	80.34.23	26	95.1 (1.77)
L32	canal	WCA-3A	26.08.83	80.40.69	25	85.3 (1.96)
L12	canal	WCA-3B	25.45.70	80.40.40	25	80.5 (1.88)
L33	canal	WCA-3B	25.56.52	80.26.37	12	88.2 (5.82)
L34	marsh	ENP	25.41.42	80.45.61	19	86.4 (3.50)
L35	marsh	ENP	25.41.32	80.45.70	12	86.0 (5.73)

Table 1 Collection information for spotted sunfish in WCA-3A and -3B, and ENP. N indicates number of specimens analysed and SL indicates mean standard length (SE). Sites are shown in Fig. 1

Fig. 1 Map of the regions sampled in the Florida Everglades showing the locations of the sample sites. The location of the sampling area is indicated in black on the Florida State map.



of significant interannual variation in allozyme or microsatellite allele frequencies in these two samples (Trexler *et al.* 2001).

We estimated the density of spotted sunfish from the marsh habitat adjacent to one of our marsh sampling sites in ENP (L35) from a long-term fish-sampling programme conducted by personnel of the park. These data were obtained by use of a 1-m^2 throw trap and approximately 105 samples were collected each year between 1978 and 1997. Fish were collected in February, April, July, October and December (for details of the sampling design see Loftus & Eklund 1994; Trexler *et al.* 2001). Juveniles of four

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Enzyme	EC no.	Ν	Р	$H_{\rm O}$
Aconitase hydratase (Acoh)	4.2.1.3	3	1.0	0.52
Aspartate aminotransferase (Aat-1)	2.6.1.1	1	0.0	0.00
Aspartate aminotransferase (Aat-2)		2	1.0	0.27
Glucose dehydrogenase (Gcdh)	1.1.1.118	3	1.0	0.09
Isocitrate dehydrogenase (Idh-1)	1.1.1.42	2	1.0	0.17
L-Lactate dehydrogenase (Ldh-2)	1.1.1.27	2	1.0	0.00
Malate dehydrogenase (Mdh-2)	1.1.1.37	1	0.0	0.00
Malate dehydrogenase (Mdh-3)		1	0.0	0.00
Mannose-6-phosphate isomerase (Mpi-1)	5.3.1.8	1	0.0	0.00
Mannose-6-phosphate isomerase (Mpi-2)		1	0.0	0.00
Phosphoglucomutase (Pgm-1)	5.4.2.2	4	1.0	0.39
Phosphoglucomutase (Pgm-2)		1	0.0	0.00
Phosphogluconate dehydrogenase (Pgdh-1)	1.1.1.14	1	0.0	0.00
Phosphogluconate dehydrogenase (Pgdh-2)		1	0.0	0.00
Peptidase val-leu (Pep)	3.4	4	1.0	0.04
Totals			0.47	0.10

Table 2 Allozymes scored for allelic variation in spotted sunfish. E.C. no. refers to Enzyme Commission number; Pep is a generic term. *N* indicates the number of alleles observed at each locus, *P* indicates the presence or absence of polymorphism in at least 1% of the individuals at each locus, and $H_{\rm O}$ is the observed heterozygosity for each locus

species of sunfish found in the Everglades are not easily distinguishable (these are *L. punctatus, L. marginatus, L. microlophus,* and *L. macrochirus*), though only spotted sunfish and *L. marginatus* are abundant. We summed the counts of spotted sunfish and unidentified juvenile sunfish to yield a generous estimate of the historical density of spotted sunfish at the sampling site prior to our collection for genetic analysis.

Allozyme analysis

We used starch-gel electrophoresis to document patterns of allozymic variation in spotted sunfish. Whole-tissue extracts were prepared for electrophoresis by homogenization of tissues in approximately 500 µL of grinding buffer (0.025 M Tris pH 7.0, 0.025 M sucrose, 0.005 M mercaptoethanol). We pooled eye, liver and soma clips; preliminary analyses indicated no tissue-specific expression of the proteins, justifying this approach. We screened 32 proteinencoding loci with horizontal starch-gel electrophoresis before selecting 15 resolvable loci to score on all individuals. Of these, seven loci were polymorphic (Table 2). We followed standard techniques described in Selander *et al.* (1971) and Murphy *et al.* (1996), with 11% (w/v) starch gels. All allozymes were resolved using a Tris-citrate EDTA buffer at pH 8.0 (Ayala *et al.* 1972).

Microsatellite DNA analysis

We surveyed nine potential microsatellite-locus primer pairs in spotted sunfish that had been developed in closely related species, *Lepomis auritus* (DeWoody *et al.* 1998) and *Lepomis macrochirus* (Colbourne *et al.* 1996). From these nine loci, we identified five polymorphic loci that amplified reliably (Table 3). Whole genomic DNA was isolated from muscle tissue by standard phenol-chloroform DNA extraction methods (Hoelzel & Green 1992). The microsatellite loci were amplified using polymerase chain reaction (PCR) in 15 μ L volumes, containing 1× buffer, 25 μ M MgCl2, 250 µм dNTP, 5 U/µL Taq DNA polymerase (Promega), and a $5 \,\mu\text{M}$ primer set. One primer of each pair was end-labelled with a fluorescent dye (6-FAM, NED or HEX; Applied Biosystems). The thermal cycling parameters, modified from Colbourne et al. (1996), were as follows: an initial 1 minute denaturation at 94 °C, followed by 45 cycles of 30 s at 94 °C, 30 s at 55 °C, and 30 s at 72 °C, and a final 10 minute extension at 72 °C. We electrophoresed the amplified samples in 2.5% agarose gel to determine the presence or absence of a product. The amplified products were electrophoresed in 5% denaturing polyacrylamide gels on an ABI 377 automated DNA sequencer. The alleles were sized with respect to electrophoretic mobility compared to a ROX 350 standard and the genotypes were assigned using the GENESCAN (ABI) and GENOTYPER software packages.

Statistical methods

We examined patterns of allelic and genotypic diversity in both allozymes and microsatellites. We calculated goodnessof-fit to Hardy–Weinberg expectations at each allozyme and microsatellite locus within each site, and in global tests across loci and across sites, using GENEPOP 3.2 (Raymond & Rousset 1995). When only two or three alleles were observed, we used the complete enumeration method of Louis & Dempster (1987), while we used the Markov chain method to calculate exact *P*-values for loci that exhibited more than three alleles (Guo & Thompson 1992). Linkage disequilibrium was tested for each pair of allozyme and microsatellite loci within each site using GENEPOP 3.2

Locus	Dye Marker	Locus Sequence Primer Sequences	Ν	$H_{\rm O}$
RB7*	6FAM	not reported F: 5' – GTGCTAATAAAGGCTACTGTC R: 5' – TGTTCCCTTAATTGTTTTGA	17	0.816
RB20*	HEX	not reported F: 5' – ggtctactggtaaatgaggg R: 5' – gttgggctgtcgagagtaaaaa	15	0.797
Lma 21**	HEX	(TC) ₁₉ (AC) ₁₁ F: 5' – cageteaatagttetgteagg R: 5' – actactgetgaagatattgtag	14	0.683
Lma 29**	6FAM	(gt) ₃₀ F: 5' – ccctgttacttgtgtattc R: 5' – attcagaggcaagcattatc	17	0.731
Lma 87**	NED	(ac) ₁₅ (a) ₅ F: 5' – atgacacagactcaccatgc R: 5' – ctcctgcccataaatcagac	14	0.566

Table 3 Primer sequences for spotted sunfish DNA microsatellites. Microsatellite fragments were sized based on migration relative to an ROX 350 standard. N indicates the number of alleles observed at each locus, and $H_{\rm O}$ is the observed heterozygosity for each locus. All samples were polymorphic for these loci

*DeWoody et al. 1998.

**Colbourne et al. 1996.

(Raymond & Rousset 1995). We used F-STAT (Goudet 2001) to estimate gene diversity and allelic richness separately for allozymes and microsatellites. We also tested for differences in gene diversity and allelic richness between allozymes and microsatellites using a Mann–Whitney's *U*-test with loci as replicates.

We hypothesized that sunfish from marsh sites that experienced seasonal dry-down events that varied in severity might experience significant reductions in their effective population size. The most recent severe drought occurred in 1989-90, when all of ENP and WCA-3B, and most of WCA-3A were dry. Thus, we examined the possible impact of demographic changes from hydrologic fluctuations or habitat fragmentation on allozyme and microsatellite diversity. We estimated deviations from expected heterozygosity and tested for evidence of recent population bottlenecks with BOTTLENECK (Cornuet & Luikart 1996; Piry et al. 1999). This program estimated heterozygosity excesses and deficiencies for each locus, and for each site, and tested significance using a one-tailed Wilcoxon test based on ranks. We used the infinite alleles model for analysis of the allozyme data, and both the stepwise and two-phase mutation models for analysis of microsatellites. In addition, an assignment test was performed according to the direct method (Rannala & Mountain 1997) with the leave-one-out option (GENECLASS software package; Cornuet et al. 1999). This permitted us to determine the likelihood that each individual's allozyme and microsatellite multilocus genotype belong to the site and /or region from which it was sampled (Paetkau et al. 1995). In order to increase the number of loci and thus the power of the population bottleneck and the assignment tests, allozyme

and microsatellite data were combined, and the above analyses were repeated. If local or regional sites are stable and genetically distinct, then recent migrants to a region or a local site might be detected and the source of these migrants could be identified.

To test for isolation by distance and equilibrium of gene flow and drift, we calculated pairwise estimates of F_{ST} and R_{ST} for allozyme and microsatellite loci, respectively (Slatkin 1993; Hutchison & Templeton 1999). We tested the significance of pairwise correlations between these values and the linear geographical distance separating sites using the Mantel permutation procedure (Mantel 1967) with the Spearman's rank correlation coefficient as the test statistic (GENEPOP 3.2; Raymond & Rousset 1995). Hydrographical distance between sites and linear geographical distance between sites were functionally equal because the Everglades is an expansive river-like wetland (i.e. with few exceptions, fish can move in all directions throughout much of the system). The patterns of the scatterplots were analysed according to Hutchison & Templeton (1999). Residual values from these analyses were classified as within or among regions, or within or among site types (canal or marsh). These values were analysed to test for regional division or site type patterns.

We used F-STAT (Goudet 2001) to compare gene diversity and allelic richness separately for allozymes and microsatellites among regions and between marsh and canal sites. We used Weir & Cockerham's (1984) co-ancestry method to hierarchically partition the total observed genetic variation for the allozyme and microsatellite data. These partitions were attributable to variation among individuals within sites (individuals collected in areas < 1 km²) relative to total diversity within their site (θ_{IS}), among sites within water-management units relative to the total diversity in that unit (θ_{SP}) , and among water-management units (regions) relative to the total genetic diversity (θ_{ST}). This provided estimates of Wright's hierarchical F-statistic of population subdivision based on a random-effects sampling model (Weir 1990). The regions correspond to an area enclosed by levees and canals: WCA-3A, WCA-3B, and ENP (Fig. 1). F-statistic estimates were also calculated separately among all canal sites and among all marsh sites to compare genetic structure estimates between deep-water refuge sites and sites that may be subject to frequent extinction and colonization. The Genetical Data Analysis program (GDA 1.0; Lewis & Zaykin 1997) was used to estimate hierarchical F-statistics following Weir & Cockerham (1984), with bootstrapping across loci to estimate 95% confidence intervals.

To make statistical comparisons between the populationgenetic patterns detected from allozymes and microsatellites, Weir & Cockerham's (1984) co-ancestry method was employed for both markers. Simulations performed by Ruzzante (1998) suggested that F_{ST} and R_{ST} perform similarly if mutation has not had enough time to affect allelic divergence of microsatellite loci. We tested for correlations of pairwise F_{ST} estimates between sites for allozymes and microsatellites with the Mantel permutation procedure performed as above.

Counts of sunfish collected in 1-m² samples deviated noticeably from a normal distribution, with many traps yielding zero specimens. Using SAS statistical software, we fit an overdispersed-Poisson linear model with a log linking function to provide estimates of annual density with asymmetrical confidence intervals lower bounded at zero (Littell *et al.* 1996).

Results

Hardy–Weinberg equilibrium and genetic diversity

Genotypic frequencies of allozymes were more consistent with Hardy-Weinberg predictions than genotypic frequencies of microsatellites. The genotypic frequencies of the allozyme loci were generally consistent with expectations of Hardy-Weinberg equilibrium; testing across loci within each site, we found only one case that did not conform to Hardy-Weinberg equilibrium expectations (L1). Additionally, we found evidence of heterozygote deficiency (f < 0) at 3 of the 7 polymorphic loci examined for that site (Table 4). Several of the microsatellite loci did not conform to Hardy-Weinberg predictions. Testing within each site indicated that genotypic frequencies within three of the sites were inconsistent with Hardy-Weinberg predictions (L1, L11, and L32). Additionally, there was significant heterozygote deficiency detected for some loci within those sites, though different loci were deficient in heterozygotes among sites (Table 5). There was some evidence for linkage disequilibrium in allozymes (*Gcdh & Pep*) and in microsatellites (Lma29 & Lma87).

We observed moderate variation at the 7 polymorphic allozymes surveyed (7 of the 15 loci examined were polymorphic = 46.7%, average heterozygosity of loci that were polymorphic = 0.25; Table 2). The number of alleles per polymorphic locus across all sites ranged from 2 (*Ldh*, *Idh*-1) to 4 (*Pgm*-1 and *Pep*). We observed considerable variation in the five microsatellite loci studied (Table 5). The number of alleles per locus ranged from 14 (Lma21 and Lma87) to 17 (RB7 and Lma29; Table 5). The observed heterozygosity, averaged over all sites, ranged from 0.57 (Lma87) to 0.82 (RB7) (Table 3).

Compared to allozymes, microsatellites had a greater average gene diversity and allelic richness (Mann-Whitney's U = 75.00; P < 0.001). Matrices of allozyme and microsatellite pairwise $F_{\rm ST}$ estimates were significantly correlated (Mantel test, r = 0.571, P = 0.009), indicating concordance between the two types of molecular markers.

Recent bottleneck test

There were no indications that any of the studied sites experienced a recent, significant bottleneck. There was one site (L1) where the number of allozyme loci showing heterozygote excess was significantly greater than that expected from the allele diversity. Under the expectations of both models of mutation, there were no sites where the number of microsatellite loci showing heterozygote excesses was significantly greater than expected from the allele diversity. After applying a Bonferroni correction for multiple tests, none of these observations could be considered statistically significant ($\alpha = 0.05$).

Assignment test

The assignment test performed with the allozyme data did not reliably assign individuals to their site or region of sampling origin. Only 57 of 231 (24.7%) were correctly assigned to the site from which they were sampled, and 80 of 231 (34.6%) correctly assigned to WCA-3A, WCA-3B or ENP. Following King et al. (2001), we found that the inclusion of all allozyme loci produced the greatest assignment accuracy. The assignment test performed with the microsatellite data did not reliably assign individuals to their site or region of sampling origin. Only 48 of 283 (17.0%) were correctly assigned to the site from which they were sampled, and 163 of 283 (57.6%) were correctly assigned to WCA-3A, WCA-3B or ENP. The inclusion and exclusion of microsatellite loci did not alter the assignment accuracy. Similarly, the assignment test performed with a pooled data set of allozymes and microsatellites **Table 4** Descriptive statistics for the 7 polymorphic allozyme loci from collections spotted sunfish at 11 sites in the southern Everglades. Number of samples genotyped (*N*), number of alleles per locus (*A*), gene diversity (H_E), allelic richness (*R*), inbreeding coefficient (*f*), and conformity to Hardy-Weinberg Equilibrium (HW)

Collection		Locus							
Site		Acon	Aat-2	Gcdh	Idh-1	Ldh-2	Pgm-1	Pep	across loci
L1	N A	23 2	23 2	23 3	13 2	23 2	16 3	23 3	
	$H_{\rm E}$	0.476	0.231	0.565	0.269	0.474	0.283	0.482	
	R	2.000	1.992	2.992	2.000	2.000	2.932	2.777	
	f	-0.004	-0.128	0.692	-0.143	1.000	-0.103	0.730	0.402
	HW	1.000	1.000	0.000	1.000	0.000	1.000	0.000	0.000
L3-	Ν	25	25	25	25	25	25	25	
	Α	3	2	2	2	1	3	2	
	$H_{\rm E}$	0.588	0.302	0.040	0.150	0.000	0.223	0.040	
	R	2.993	1.999	1.480	1.935	1.000	2.735	1.480	
	f	-0.089	0.072	0.000	-0.067	_	-0.079	0.000	-0.043
	HW	0.376	1.000	—	1.000	_	1.000	—	0.982
L11	Ν	25	25	25	25	25	25	25	
	A	3	2	1	2	1	3	1	
	$H_{\rm E}$	0.538	0.327	0.000	0.150	0.000	0.279	0.000	
	R	2.969	1.999	1.000	1.935	1.000	2.473	1.000	
	f	0.034	0.020	NA	-0.067	_	-0.146	NA	-0.020
	HW	0.886	1.000	_	1.000	_	1.000	_	1.000
L12	Ν	25	25	25	25	25	25	25	
	A	3	2	2	2	1	3	1	
	$H_{\rm E}$	0.587	0.245	0.078	0.078	0.000	0.681	0.000	
	R	2.986	1.993	1.735	1.735	1.000	3.000	1.000	
	f LIM	-0.295	-0.143	-0.021	-0.021	—	0.119	NA	-0.078
	ПМ	0.031	1.000	1.000	1.000	_	0.520	_	0.012
L21	N	14	14	14	14	14	14	14	
	A	3	2	2	2	1	3	2	
	H _E	0.527	0.308	0.137	0.253	0.000	0.467	0.071	
	K	2.857	2.000	1.984	2.000	1.000	2.857	1.857	0.020
	f LIMA	0.323	0.304	-0.040	-0.130	_	-0.376	0.000	0.028
		0.236	0.540	1.000	1.000	_	0.400	_	0.772
L23	N	25	25	25	25	25	25	25	
	A	3	2	2	2	1	4	3	
	H _E	0.423	0.245	0.115	0.115	0.000	0.603	0.079	
	K	2.480	1.993	1.867	1.867	1.000	3.466	1.960	0.010
	J LIM	0.053	-0.143	-0.043	-0.043	_	0.006	-0.011	-0.013
		1.000	1.000	1.000	1.000	_	0.047	1.000	1.000
L25	N	23	26	26	26	26	21	26	
	A	2	2	2	2	1	3	1	
	H _E	0.451	0.265	0.111	0.111	0.000	0.479	0.000	
	ĸ	2.000	1.996	1.852	1.852	1.000	2.822	1.000 NIA	0 1 2 9
	J HW	0.132	-0.163	-0.042	-0.042	_	-0.393	NA _	-0.128
1.20	N	22	25	25	25	25	25	24	01007
L32	IN A	23	25	25 1	25	25 1	25	24	
	И_ Н_	0.611	0 245	0.000	0 150	0.000	$\frac{2}{0.040}$	0 082	
	R	2 992	1 993	1 000	1 935	1 000	1 480	2 000	
	f	0.074	-0.143	NA	0.067		0.000	-0.011	0.000
	HW	0.158	1.000	_	1.000	_	_	1.000	0.884

Table 4	Continued
	Commen

Callection		Locus							
Site		Acon	Aat-2	Gcdh	Idh-1	Ldh-2	Pgm-1	Рер	across loci
L33	Ν	12	12	12	12	12	12	12	
	Α	3	2	2	2	1	4	1	
	$H_{\rm F}$	0.591	0.159	0.288	0.348	0.000	0.420	0.000	
	R	3.000	2.000	2.000	2.000	1.000	4.000	1.000	
	f	0.295	-0.048	-0.158	0.283	_	-0.189	NA	0.078
	HW	0.169	1.000	1.000	0.403	—	1.000	—	0.865
L34	Ν	17	19	19	17	19	18	19	
	Α	3	2	2	2	1	1	1	
	$H_{\rm E}$	0.465	0.430	0.102	0.114	0.000	0.000	0.000	
	R	2.706	2.000	1.871	1.920	1.000	1.000	1.000	
	f	-0.265	0.633	-0.029	-0.032	_	NA	NA	0.128
	HW	0.082	0.012	1.000	1.000	-	—	-	0.086
L35	Ν	12	12	12	12	12	12	12	
	Α	2	2	2	2	1	3	1	
	$H_{\rm E}$	0.462	0.288	0.167	0.288	0.000	0.595	0.000	
	R^{-}	2.000	2.000	2.000	2.000	1.000	3.000	1.000	
	f	-0.082	-0.158	1.000	-0.158	_	0.019	NA	0.027
	HW	1.000	1.000	0.043	1.000	_	1.000	_	0.792
Total all	Ν	179	186	186	174	184	173	185	
sites	Α	3	2	3	2	2	4	4	
	$H_{\rm F}$	0.520	0.277	0.146	0.184	0.043	0.367	0.069	
	R	2.808	1.988	2.223	1.994	1.994	2.974	1.805	
	f	0.006	0.026	0.396	-0.034	1.000	0.078	0.556	0.126
	HW	0.342	0.975	0.007	1.000	1.000	0.965	0.003	0.000

Table 5 Descriptive statistics for five microsatellite loci in collections of spotted sunfish from 11 sites in the Everglades. Number of samples
genotyped (N), number of alleles per locus (A), gene diversity (H _E), allelic richness (R), inbreeding coefficient (f), and conformity to Hardy-
Weinberg equilibrium are reported

		Locus					
Collection Site		RB7	Lma21	Lma29	RB20	Lma87	All loci
L1	Ν	22	23	23	20	23	
	Α	13	7	11	11	5	
	$H_{\rm F}$	0.918	0.791	0.848	0.850	0.610	
	R	9.025	5.561	7.472	7.152	4.004	
	f	0.257	0.120	0.282	0.176	0.073	0.190
	HW	0.000	0.001	0.000	0.000	0.208	0.000
L3	Ν	13	13	13	13	13	
	Α	10	5	8	12	6	
	$H_{ m E}$	0.897	0.718	0.849	0.923	0.785	
	R	8.209	4.463	7.104	9.551	5.416	
	f	-0.029	-0.393	-0.177	0	0.020	-0.106
	HW	0.854	0.024	0.906	0.574	0.597	0.432
L11	Ν	31	31	31	30	31	
	Α	13	8	12	12	7	
	$H_{\rm F}$	0.887	0.711	0.891	0.923	0.751	
	R	8.111	4.659	8.344	9.059	4.967	
	f	-0.019	0.092	0.095	0.350	0.485	0.197
	HW	0.184	0.114	0.271	0.000	0.000	0.000

Table 5 Continued

Callertian		Locus					
Site		RB7	Lma21	Lma29	RB20	Lma87	All loci
L12	Ν	45	40	45	43	45	
	Α	15	10	16	14	9	
	$H_{\rm F}$	0.908	0.765	0.885	0.909	0.743	
	R	8.887	5.967	8.436	8.725	5.273	
	f	0.045	0.150	0.071	0.027	0.073	0.071
	HW	0.117	0.006	0.213	0.252	0.519	0.018
L21	Ν	23	22	23	22	21	
	Α	12	9	11	15	8	
	$H_{\rm F}$	0.914	0.766	0.813	0.935	0.785	
	R	8.888	6.079	6.998	10.173	5.976	
	f	0.144	-0.008	-0.016	0.076	0.150	0.072
	HW	0.428	0.441	0.938	0.250	0.231	0.517
L23	Ν	12	11	8	10	12	
	Α	11	4	9	7	7	
	H_{r}	0.932	0.745	0.911	0.900	0.790	
	R	9.347	3.985	9.000	6.754	5.653	
	f	0.106	-0.098	0.039	0.444	0.194	0.143
	HW	0.131	0.026	0.735	0.001	0.331	0.002
L25	Ν	34	33	33	26	33	
	Α	14	7	15	13	6	
	$H_{\rm F}$	0.906	0.746	0.910	0.927	0.694	
	R^{E}	8.957	4.587	9.410	9.583	4.042	
	f	0.156	0.107	0.034	0.087	0.257	0.122
	HW	0.108	0.329	0.787	0.048	0.295	0.114
L32	Ν	50	43	.51	30	44	
202	A	14	9	12	14	10	
	Hn	0.905	0.744	0.900	0.918	0.754	
	R	8.719	4.792	8.329	9.387	5.134	
	f	0.116	0.000	0.390	0.129	0.187	0.169
	HW	0.009	0.009	0.000	0.094	0.006	0.000
L33	Ν	21	20	19	19	21	
200	A	11	5	9	14	6	
	H ₋	0.910	0.761	0.865	0.920	0.733	
	R	8 646	4.325	7 029	9.357	5 172	
	f	0.005	0.737	0 514	-0.087	0.545	0.317
	HW	0.157	0.000	0.000	0.888	0.000	0.000
L34	Ν	14	12	14	14	14	
	A	9	4	12	12	8	
	H_{r}	0.860	0.705	0.923	0.912	0.810	
	R	7.568	3.666	9.426	9.124	6.505	
	f	0.003	-0.301	0.149	0.139	0.207	0.053
	HW	0.491	0.354	0.361	0.195	0.253	0.317
L35	Ν	12	12	10	12	11	
	A	9	4	7	11	5	
	H_{r}	0.890	0.784	0.867	0.917	0.800	
	R	7.898	3.997	6.730	9.225	4.920	
	f	0.157	0.362	-0.154	0.000	-0.023	0.064
	HW	0.097	0.163	0.971	0.571	0.739	0.436
Total all	Ν	305	279	294	257	289	
Populations	A	17	14	17	15	14	
r	Hr	0.902	0.749	0.878	0.913	0.750	
	R	8.78	5.20	8.51	9.11	5.30	
	f	0.131	0.091	0.161	0.152	0.226	0.143
	HW	0.000	0.000	0.000	0.000	0.000	0.000

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Fig. 2 Relationship between geographical distance and (A) θ_{ST} for the allozyme data (B) θ_{ST} for the microsatellite data, and (C) R_{ST} for the microsatellite data.

did not reliably assign individuals to their site or region of sampling origin. Only 58 of 231 (25.1%) were correctly assigned to the site from which they were sampled, and 114 of 231 (49.4%) were correctly assigned to WCA-3A, WCA-3B, or ENP. This suggests that there is not a unique multilocus genotypic distribution for the allozyme or microsatellite loci that described a site or region. This could result from routine mixing within and/or between the regions. The assignment test assumes Hardy-Weinberg equilibrium, which was partially violated by the microsatellite data.

Isolation by distance

We found no correlations between pairwise estimates of geographical distance and $F_{\rm ST}$ from allozymes or $F_{\rm ST}$ and $R_{\rm ST}$ from microsatellite loci, further indicating a lack of spatially explicit genetic structure (Fig. 2). Furthermore, residual values calculated from these regressions did not reveal any significant patterns of deviation associated with region or site type. The overall pattern of the scatterplots was consistent between allozymes and microsatellites. **Table 6** Patterns of genetic diversity in spotted sunfish collected from the southern Everglades. Estimates reported for AAll Sites@ were hierarchically organized as sites within regions, and within each region or habitat type as listed. θ_{SP} indicates the genetic variance among sites within regions. θ_{PT} indicates the genetic variance in among samples relative to the total study

	Microsate	llite data	Allozyme data		
	$\theta_{\rm SP}$	θ_{PT}	θ_{SP}	θ_{PT}	
All Sites Canal Marsh	0.0052* NA NA	-0.0001 0.0021 0.0229*	0.0745* NA NA	-0.0068 0.0485 0.1446*	

*indicates 95% confidence interval does not cross zero.

Hierarchical genetic diversity

Gene diversity and allelic richness were not significantly different between WCA-3A, WCA-3B and ENP (P > 0.05) for allozymes and microsatellites. They were not significantly different between marsh and canal sites for allozymes (P > 0.05). However, gene diversity and allelic richness were significantly different between marsh and canal sites for microsatellites (gene diversity; P = 0.022: allelic richness; P = 0.009). Marsh sites had, on average, less gene diversity and allelic richness than canal sites.

Allozymes and microsatellites revealed consistent and significant genetic structure among sites within regions (Table 6). When considered separately, sites within WCA-3A displayed significant genetic structure, but sites in WCA-3B and ENP did not (i.e. bootstrap confidence intervals cross zero). The F-statistic estimates were an order of magnitude larger for allozymes than for microsatellites. This is consistent with expectations of *F*-statistic estimates calculated from low to moderately variable loci compared to highly variable loci (reviewed in Hedrick 1999). Canal sites were not genetically differentiated, whereas marsh sites were significantly heterogeneous (confidence interval of $\theta_{PT marsh}$ did not cross zero). Based on a *t*-test re-sampling across loci with 1000 replicates, marsh sites were more heterogeneous than canal sites (H_{O} : $\theta_{PT \text{ canals}} = \theta_{PT \text{ marsh}}$; P = 0.115 for microsatelites and P = 0.109 for allozymes; Re-sampling test combining both data sets P = 0.08).

Spotted sunfish population dynamics

Spotted sunfish density in the marsh habitat near our ENP sampling sites fluctuated between approximately 0.5/m² to indistinguishable from zero over the 18 years prior to our study (Fig. 3a). No specimens of spotted sunfish (or of indistinguishable juveniles of the co-occurring congeners) were collected in 1988 or 1989, and very few specimens were collected in 1990. These latter low-density years overlapped with a relatively extreme 2-year drought event



Fig. 3 Long-term record of density of spotted sunfish in marsh habitat adjacent to site L35 in ENP. (A) The average density of spotted sunfish (including all indistinguishable juvenile sunfish) in a marsh habitat with 95% confidence intervals estimated on an annual basis from 1978 to 1997. Note that no spotted sunfish were collected in 1988 or 1989, and very few in 1990. Asterisk indicates year when sampling was missed for 2 of the 5 months. (B) Annual minimum water depth at the collection site between 1978 and 1997.

in 1989 and 1990. A caveat is required for the low estimate in 1988; that is the only year of the 2-year record where sampling events were missed but water levels were above the ground surface; no samples were collected in April or May of that year.

Discussion

The spotted sunfish data were not consistent with use of regional-scale island models (at the spatial scale of water management units or greater) to describe the population structure for this species. The data suggested that spotted sunfish populations could be described as metapopulations that are not at equilibrium between gene flow and genetic drift. We propose that local-scale population dynamics have a significant effect on the population genetic structure of spotted sunfish in the Everglades. We found no evidence of additional regional-scale variation. There are three possible sources of the lack of regional structure: historical population structure has been lost because of the homogenizing effects of canals as conduits for dispersal; the ecosystem historically had high gene flow and the canals and levees have not altered that situation; or, no genetic structure existed historically and, though the modern management system now limits gene flow, inadequate time has passed (approximately 20 generations of spotted sunfish) for the large regional populations to differentiate. Below, we elaborate on these conclusions.

Hardy-Weinberg equilibrium

Some sites were deficient in the frequency of heterozygotes, especially for microsatellite loci. Heterozygote deficiencies can arise from inbreeding, underdominant selection, mixing of genetically differentiated populations at a site (spatial Wahlund effect), or mixing of cohorts (temporal Wahlund effect) (Hartl & Clark 1997). The concordance we observed between allozyme and microsatellite markers (pairwise $F_{\rm ST}$ Mantel test) suggests that gene flow and genetic drift were the major causes of any observed patterns of population differentiation (cf., Estoup *et al.* 1998; Ross *et al.* 1999). The absence of widespread patterns of heterozygote deficiency suggests that Wahlund effects are not principal factors in the genetic structuring of local populations of spotted sunfish.

The coefficient of variation (CV) of body size of individuals collected from site L1 was much greater than that of any other site sampled ($CV_{1.1} = 8.91$; $CV_{OTHER SITES} = 2.76$). The average size was not significantly different among sites with the exception of site L23 (Table 1). Although definitive discrimination between spatial and temporal Wahlund effects is beyond the power of our data, we pooled individuals into 1.0 cm size classes across all sites (6 size classes total, sample sizes within each class ranged from 14 to 39). We were unable to create size-based cohorts within sites because of low sample size. We re-tested for consistency with Hardy-Weinberg expectations within size classes for the microsatellites, and found only one locus (M87) that significantly deviated from expectations and only one of the cohorts was inconsistent with predicted frequencies (size class 6, 10.0 + cm, a mixture of larger/older classes). This suggests that temporal Wahlund effects should be considered in future studies and highlights the possibility that temporal variation and genetic differentiation between sites may be transient in time (Tessier & Bernatchez 1999). Genetic structure could be modified each generation by the addition of new cohorts and the mixing of old cohorts, depending on the connectivity of sites throughout the Everglades marsh. Thus, unique genetic signatures within sites may be masked by temporal variation (Whitlock & McCauley

1990; David *et al.* 1997); however, this does not preclude the possible effects of spatial mixing. It is interesting to note that pooling among sites within regions or among regions did not improve the fit of the data to Hardy-Weinberg expectations.

Patterns of genetic diversity

Allozyme and microsatellite loci differed in the amount of genetic diversity they showed, and in the patterns of gene diversity and allelic richness among sites. Highly variable microsatellite loci are likely more sensitive to these measures of genetic differentiation (Hedrick 1999). Both markers showed no evidence for recent population bottlenecks, and assignment tests did not reliably assign individuals to the site or region from which they were sampled. The assignment test failed to identify a unique genetic signal for regions and sites within regions. Thus, colonization of the marsh habitat after the dry season is not likely the result of a very small number of individuals that survive through the dry season in a deep-water refuge or colonize from canals to populate the site through *in situ* reproduction. Rather, the data indicate that a moderate to large immigrant population of spotted sunfish probably colonizes marsh sites. Also, a recent severe drought in 1989-90 did not cause a detectable bottleneck in those marsh sites, or subsequent colonization of the marsh sites from canal sites has masked the effects of the bottleneck.

We observed no evidence for isolation by distance in either allozymes or microsatellites, conforming to Hutchison & Templeton's (1999) case II. This pattern suggests that populations of spotted sunfish are not in equilibrium with respect to gene flow and genetic drift, possibly because of a recent colonization event from a relatively homogeneous source population. This conclusion is supported by the hierarchical population structure analyses, which did not reveal population structure for allozymes or microsatellites at the regional scale. Small but significant population structure and genetic differences were detected at the local level, which were likely the result of extinction and colonization effects at local spatial scales.

Local scale extinction-colonization dynamics of fish are tightly coupled with the natural hydrologic cycles of the Everglades (Trexler *et al.* 2001). We illustrated that the pattern with a 20-year record of spotted sunfish density in the marshes adjacent to one of our sites used for sampling genetic variation. Approximately 6 years prior to our study, the density of spotted sunfish dropped very low (indistinguishable from zero) for 1 to 2 years (1989 and 1990, we exclude the 1988 result because of low sample size). In 1990, alligator ponds in the vicinity of this marshsampling site dried, in addition to the marsh, because ground water depths dropped below 60 cm from the ground surface (Fig. 3b). We believe that this pattern of population dynamics is linked to our finding of modest population structure among marsh sites, but not among canal sites for spotted sunfish. This pattern is present even though the average distance between the marsh sites (16.8 km) was less than (P = 0.104) the average distance between canal sites (30.2 km). Such a pattern of population genetic structure is expected when a group of recently colonized populations is compared to populations that are not subject to extinction and colonization turnover (Whitlock & McCauley 1990). Additionally, the low F-statistics and small difference between canal and marsh F-statistics suggests that spotted sunfish metapopulations are intermediate along a gradient between propagule pool and migrant pool patterns (McCauley et al. 1995). One might also expect this result if 'routine' migration has diluted the effects of a population crash on patterns of genetic variation, possibly as in our study where the crash was noted in 1990, six or more years prior to our collections.

Our data suggest that canal sites yield a homogeneous pool of colonists for marsh sites that dry routinely. We hypothesize that marsh sites are colonized by a modest number of individuals from such sources representing a subset of the total genetic diversity present there, but not diminishing genetic variation enough to be detected statistically as a 'bottleneck'. We propose that these colonization events create a modest genetic signature in the resulting marsh populations that may be short lived. Thus, a continent-island (canal-marsh) population structure best describes spotted sunfish genetics in the Everglades.

Genetic variation of Everglades spotted sunfish may not be at equilibrium at the regional level, in addition to the local dynamics already discussed. We do not know the historical population structure of sunfishes in the Everglades, but the regional habitat structure we sampled is a relatively recent addition to the Everglades ecosystem. The Tamiami Canal, separating WCA-3A and ENP, was originally constructed in 1926, and WCA-3A was enclosed by levees in 1963, thus the present regional habitat structure has been in place for 35-80 years. A minimum generation time for spotted sunfish is 2 years, so that 18–40 generations have passed since the regional structure was superimposed on the ecosystem. If the historical population was unstructured and the canals and levees form barriers to gene flow, the contemporary populations may not be at equilibrium with these changes. Genetic differentiation by drift, even with complete isolation (which is only possible for WCA-3B), is unlikely to have occurred in this short time without strong limits to dispersal and severe population bottlenecks (e.g. Arnaud et al. 2001).

The Everglades' natural hydropattern has been disrupted by recent water management practices. Regions of the Everglades that were historically flooded only a fraction of the year, are now flooded for prolonged periods of time, while other areas are dry much longer each year than they were historically (Fennema *et al.* 1994). Also, historical patterns of sheet flow (flow of water across the wetland landscape driven by rainfall events) have largely been replaced with pulsed flow tied to management needs, resulting in lacustrine (southern WCA-3A) or stagnant conditions (WCA-3B) at some locations (Light & Dineen 1994; DeAngelis *et al.* 1997). Hydrologic patterns may affect the efficiency of colonization and exploitation of marsh habitats that become seasonally available to aquatic species in the Everglades. Our data suggest that patterns of periodic dry-down may create recurrent mixing events for spotted sunfish and, in this way, water management decisions may shape patterns of genetic diversity of this species in the contemporary Everglades ecosystem.

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References

- Arnaud JF, Madec L, Guiller A, Bellido A (2001) Spatial analysis of allozyme and microsatellite DNA polymorphisms in the land snail *Helix aspersa* (Gastropoda: Helicidae). *Molecular Ecology*, 10, 1563–1576.
- Ayala JW, Powell JR, Tracey ML, Mourao CA, Perez-Salas S (1972) Enzyme variability in the *Drosophila willistoni* group. IV. Genic variation in natural populations of *Drosophila willistoni*. *Genetics*, **70**, 113–139.
- Blake NM (1980) Land into Water Water into Land. A History of Water Management in Florida. University Presses of Florida, Tallahassee.
- Colbourne JK, Neff BD, Wright JM, Gross MR (1996) DNA fingerprinting of bluegill sunfish (*Lepomis macrochirus*) using (GT)_n microsatellites and its potential for assessment of mating success. *Canadian Journal of Fisheries and Aquatic Sciences*, **53**, 342– 349.
- Cornuet J-M, Luikart G (1996) Description and power analysis of two tests for detecting recent population bottlenecks from allele frequency data. *Genetics*, **144**, 2001–2014.
- Cornuet J-M, Piry S, Luikart G, Estoup A, Solignac M (1999) New methods employing multilocus genotypes to select or exclude populations as origins of individuals. *Genetics*, **153**, 1989–2000.
- David P, Perdieu M-A, Pernot A-F, Jarne P (1997) Fine-grained spatial and temporal population genetic structure in the marine bivalve *Spisula ovalis*. *Evolution*, **51**, 1318–1322.
- DeAngelis DL, Loftus WL, Trexler JC, Ulanowicz RE (1997) Modeling fish dynamics and effects of stress in a hydrologically
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pulsed system. *Journal of Aquatic Ecosystem Stress and Recovery*, **6**, 1–13.

- DeWoody JA, Fletcher DE, Wilkins SD, Nelson WS, Avise JC (1998) Molecular genetic dissection of spawning, parentage, and reproductive tactics in a population of redbreast sunfish, *Lepomis auritus. Evolution*, **52**, 1802–1810.
- Estoup A, Rousset F, Michalakis Y, Cornuet J-M, Adriamanga M, Guyomard R (1998) Comparative analysis of microsatellite and allozyme markers: a case study investigating microgeographic differentiation in brown trout (*Salmo trutta*) *Molecular Ecology*, **7**, 339–353.
- Fennema RJ, Neidrauer CJ, Johnson RA, MacVicar TK, Perkins WA (1994) A computer model to simulate natural Everglades hydrology. In: *Everglades, the Ecosystem and its Restoration* (eds Davis SM, Ogden JC), pp. 249–289. St Lucie Press, Boca Raton, FL.
- Goudet J (2001) FSTAT, a Program to Estimate and Test Gene Diversities and Fixation Indices Version 2.9.3. http://www.unil.ch/izea/ softwares/fstat.html.UpdatedfromGoudet.
- Guo SW, Thompson EA (1992) Performing the exact test of Hardy-Weinberg proportions for multiple alleles. *Biometrics*, **48**, 361– 372.
- Hartl DL, Clark AG (1997) Random genetic drift. In: *Principles of Population Genetics*, 3rd edn, pp. 267–313. Sinauer Associates, Inc., Sunderland, MA.
- Hedrick PW (1999) Highly variable loci and their interpretation in evolution and conservation. *Evolution*, **53**, 313–318.
- Hoelzel AR, Green A (1992) Analysis of population-level variation by sequencing PCR-amplified DNA. In: *Molecular Genetic Analysis of Populations: a Practical Approach* (ed. Hoelzel AR), pp. 159– 186. IRL Press, Oxford.
- Hutchison DW, Templeton AR (1999) Correlation of pairwise genetic and geographic distance measures: inferring the relative influences of gene flow and drift on the distribution of genetic variability. *Evolution*, **53**, 1898–1854.
- King TL, Kalinowski ST, Schill WB, Spidle AP, Lubinski BA (2001) Population structure of Atlantic salmon (*Salmo salar* L.): a rangewide perspective from microsatellite DNA variation. *Molecular Ecology*, **10**, 807–821.
- Lewis PO, Zaykin D (1997) Genetic Data Analysis: computer program for the analysis of allelic data. http://chee.unm.edu/gda/.
- Light SS, Dineen JW (1994) Water control in the Everglades: a historical perspective. In: *Everglades, the Ecosystem and its Restoration* (eds Davis SM, Ogden JC), pp. 47–84. St Lucie Press, Boca Raton, FL.
- Littell RC, Milliken GA, Stroup WW, Wolfinger RD (1996) SAS System for Mixed Models. SAS Institute Inc, Cary, NC.
- Loftus WF, Eklund AM (1994) Long-term dynamics of an Everglades small-fish assemblage. In: *Everglades, the Ecosystem and its Restoration* (eds Davis SM, Ogden JC), pp. 461–484. St Lucie Press, Boca Raton, FL.
- Loftus WF, Kushlan JA (1987) Freshwater fishes of southern Florida. *Bulletin of the Florida State Museum, Biological Science*, **31**, 147–344.
- Louis EJ, Dempster ER (1987) An exact test for Hardy-Weinberg and multiple alleles. *Biometrics*, 43, 805–811.
- Mantel N (1967) The detection of disease clustering and generalized regression approach. *Cancer Research*, 27, 209–220.
- McCauley DE, Raveill J, Antonovics J (1995) Local founding events as determinants of genetic structure in a plant metapopulation. *Heredity*, **75**, 630–636.
- Murphy RW, Sites JW Jr, Buth DG, Haufler CH (1996) Proteins:

Isozyme electrophoresis. 120. In: *Molecular Systematics*, 2nd edn (eds Hillis DM, Moritz C, Mable BK), pp. 51–120. Sinauer, Sunderland, MA.

- Nagylaki T (1998) Fixation indices in subdivided populations. *Genetics*, **148**, 1325–1332.
- Paetkau D, Calvert W, Stirling I, Strobeck C (1995) Microsatellite analysis of population structure in Canadian polar bears. *Molecular Ecology*, 4, 347–354.
- Pannell JR, Charlesworth B (1999) Neutral genetic diversity in a metapopulation with recurrent local extinction and recolonization. *Evolution*, 53, 664–676.
- Piry S, Luikart G, Cornuet JM (1999) BOTTLENECK: a computer program for detecting recent reductions in the effective population size using allele frequency data. *Journal of Heredity*, **90**, 502–503.
- Rannala B, Mountain JL (1997) Detecting immigration by using multilocus genotypes. Proceedings of the National Academy of Sciences USA, 94, 9197–9221.
- Raymond M, Rousset F (1995) GENEPOP Version 32.: population genetics software for exact tests and ecumenicism. *Journal of Heredity*, **86**, 248–249.
- Ross KG, DeWayne Shoemaker D, Krieger MJB, Deheer CJ, Keller L (1999) Assessing genetic structure with multiple classes of molecular markers: a case study involving the introduced fire ant *Solenopsis invicta*. *Molecular Biology and Evolution*, **16**, 525–543.
- Ruzzante DE (1998) A comparison of several measures of genetic distance and population structure with microsatellite data: Bias and sampling variance. *Canadian Journal of Fisheries and Aquatic Sciences*, 55, 1–14.
- Selander RK, Smith MH, Yang SY, Johnson WE, Gentry JR (1971) Biochemical polymorphism and systematics in the genus *Pero-myscus* I Variation in the old-field mouse (*Peromyscus polionotus*) stud gene VI. University Texas Publications, **7103**, 49–90.

- Slatkin M (1977) Gene flow and genetic drift in a species subject to frequent local extinctions. *Theoretical Population Biology*, **12**, 253– 262.
- Slatkin M (1993) Isolation by distance in equilibrium and nonequilibrium populations. *Evolution*, **47**, 264–279.
- Tessier N, Bernatchez L (1999) Stability of population structure and genetic diversity across generations assessed by microsatellites among sympatric populations of landlocked Atlantic salmon (*Salmo salar* L.). *Molecular Ecology*, 8, 169–179.
- Trexler JC, Loftus WF, Jordan F, Chick JH, Kandl KL, McElroy TC, Bass OL (2001) Ecological scale and its implications for freshwater fishes in the Florida Everglades. In: *The Everglades, Florida Bay, and Coral Reefs of the Florida Keys An Ecological Sourcebook* (eds Porter JW, Porter KG), pp. 153–181. CRC Press, Boca Raton, FL.
- Wade MJ, McCauley DE (1988) Extinction and recolonization: their effects on the genetic differentiation of local populations. *Evolution*, **42**, 995–1005.
- Weir BS (1990) Genetic Data Analysis. Sinauer, Sunderland, MA.
- Weir BS, Cockerham CC (1984) Estimating *F*-statistics for the analysis of population structure. *Evolution*, **38**, 1358–1370.
- Whitlock MC, McCauley DE (1990) Some population genetic consequences of colony formation and extinction: Genetic correlations within founding groups. *Evolution*, **44**, 1717–1724.

Thom McElroy, Karen Kandl, and Joel Trexler are interested in the role of population dynamics in shaping population genetic structure of aquatic organisms in fluctuating environments. This work is part of our larger project on the metacommunity dynamics of aquatic organisms in the Florida Everglades. Janette Garcia assisted with microsatellite analysis in developing her senior undergraduate thesis at Florida International University.